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Review

A Review on Mariculture Effluent: Characterization and Management Tools

Xinyan Wang ¹, Alan Cuthbertson ², Carlo Gualtieri ³  and Dongdong Shao ^{1,4,*}

¹ State Key Laboratory of Water Environment Simulation & School of Environment, Beijing Normal University, Beijing 100875, China; wangxinyanwest@126.com

² School of Science and Engineering, University of Dundee, Dundee DD1 4HN, Scotland, UK; a.j.s.cuthbertson@dundee.ac.uk

³ Department of Civil, Architectural and Environmental Engineering, University of Napoli Federico II, Corso Umberto I 40, 80138 Napoli, Italy; cagualti@unina.it

⁴ Tang Scholar, Beijing Normal University, Beijing 100875, China

* Correspondence: ddshao@bnu.edu.cn; Tel.: +86-10-5880-5053

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Abstract: While marine aquaculture, or mariculture, has been growing rapidly and globally in recent decades, many environmental concerns remain to be fully addressed to achieve its long-term goal of sustainable development. This paper aims to provide a synthesized perspective on these issues by reviewing and discussing the characterization, transport, and current modelling and management tools associated with effluents released from mariculture sites. Specifically, we examined the effluent characteristics and behavior from source-to-sink, including the composition and load of effluent discharge, its transport and transformation processes in the water column and at the seabed, and its impacts on the pelagic and benthic environments. We then focused on management-related issues, including the setting of the regulatory mixing zone, the establishment of environmental standards, monitoring measures, and modelling techniques to depict the current state-of-the-art modes in a global context. Our study shows that while substantial progress has been made in understanding the nature of the mariculture effluent, as well as in monitoring and modelling its transport and fate, the regulatory framework still lags behind in many countries where the mariculture industry is relevant. This is particularly evident in the lack of consistent criteria for the definition of regulatory mixing zones and the associated environmental standards for water quality and benthic impacts. Besides, as new predictive models are emerging quickly, their proper evaluation and validation are imperative in view of their increasing application in regulatory practices. This review is intended to provide references for advancing regulatory management of mariculture effluents, as well as for promoting sustainable mariculture development.

Keywords: mariculture; effluent discharge; effluents characterization; regulatory mixing zones; environment standards; environmental monitoring; near-field and far-field water quality models

1. Introduction

Marine aquaculture, or mariculture, is one of the fastest-growing food-producing industries in the world. Globally, the scale and production of mariculture have expanded rapidly to fulfill the protein demand for an ever-increasing population to achieve the United Nations' sustainable development goals [1,2]. Mariculture is dominated by unfed cultivation of aquatic plants, filter-feeding bivalves and mollusks, the fed cultivation of marine finfish (primarily salmonoid species), and shrimps [3]. Geographically, mariculture is distributed worldwide, with production concentrated in specific countries including China, Indonesia, India, Norway, Chile, Japan, and Scotland, amongst which China is by far the largest producer (20.65 million tons in 2019).

The potential interaction and conflict between mariculture activities and marine environments and ecosystems is a long-known issue of concern [4], with aquaculture-derived effluents one of the most critical aspects that contribute to detrimental environmental impacts. Through recurring input of feeds, fertilizers, and chemicals that are necessary to ensure the health and growth within the culture, mariculture systems generate considerable amounts of effluent that can have undesirable impacts on the environment [5,6]. Abundant loads of organic wastes, nutrients, and chemicals in the effluent are discharged into the receiving waters, and are subsequently transported in the water column and settle on the seabed, impacting the adjacent area (near-field) and beyond (far-field) [7,8]. These can cause significant environmental and ecological issues such as low levels of dissolved oxygen (i.e., hypoxia), eutrophication, heavy metal pollution, and habitat destruction [9]. Appropriate stringent regulatory measures are therefore required to manage the discharge of mariculture effluents and their impacts [10–12].

At the same time, by impacting the local aquatic environment with a given assimilative capacity, the effluent discharge also affects the ecological carrying capacity of the mariculture site, i.e., the maximum production without unacceptable impacts on ecosystems [13]. Sustainable mariculture requires the ability to balance human development goals while simultaneously sustaining marine systems to provide natural food resources and ecosystem services [14–16]. Assessment of this carrying capacity is therefore critical for the sustainability of mariculture, in which the characteristics and impacts of the associated effluent discharge constitute important elements that need to be taken into account alongside the location, type, and size of mariculture systems. In addition to these potential negative effects, some environmental and economic benefits of mariculture effluent have been recently assessed to achieve the balance between economic income and environmental loss [17]. For instance, the integrated multi-trophic aquaculture (IMTA) [18,19] is currently gaining popularity worldwide, as it spatially integrates organic waste and nutrient discharge systems such as fish cages with the culture of nutrient-extracting organisms such as mussels and algae [20]. In either case, it is necessary to elucidate the underlying nature and behavior of the effluent from its source to sink so that carrying capacity assessment and sustainable management can be properly designed and assessed.

To provide a synthesized perspective on the issue of mariculture effluent, this review focuses on the discharge characteristics, transport and transformation processes of effluent from the mariculture system, as well as the associated environmental impacts and current management measures and tools. From a global perspective, we compared environmental quality standards and regulatory policies of discharged effluent and mixing zones from a number of major mariculture countries. The monitoring and modelling techniques utilized to measure and predict impacts were reviewed to identify state-of-the-art regulatory practices. Consequently, this review was intended to provide an overview of the environmental impacts of current mariculture discharge and references for advancing regulatory management of mariculture effluent discharge, as well as promoting sustainability in future mariculture developments.

2. Types of Mariculture Systems

The majority of mariculture sites are distributed in nearshore marine waters and inter-tidal regions (e.g., mudflats) based on the habitat requirements of the breeding species. There have also been significant recent developments in the emerging area of offshore mariculture driven partly by deterioration in the nearshore environment and the advancement of offshore technologies [21]. The mariculture systems are comprised of four typical types: marine cage culture, marine pond or tank culture, marine poles, ropes, and net bags culture for mollusks and marine suspended culture for aquatic plants (detailed information on typical mariculture systems can be found in [22]) (Figure 1).

Marine cage culture is generally used for fish farming and refers to the activity of setting enclosed structures such as cages and pens near the surface in seawater bodies and cultivating fish within them [23]. Currently, such cage cultivation systems provided the vast majority of fish products

globally, alongside other cultivation systems producing other aquatic products such as shellfish and seaweeds [24]. Since the fish cages are essentially an open system where water passes freely through them and interact strongly with the environment, cage systems tend to produce wastes that are released directly into the marine environment [25]. Moreover, fish cage cultivation, unlike the cultivation of other species, requires feeding with additional concentrations of nutrients and chemicals to promote fish growth and help prevent disease, while the metabolic processes of fish also yield excessive pollutants [26]. Therefore, cage cultivation represents the main source of waste effluent discharge among all mariculture systems (even though it only constitutes about 5% of total mariculture production, Figure 1). As such, this paper focused mainly on these marine cage systems and the effluents produced.

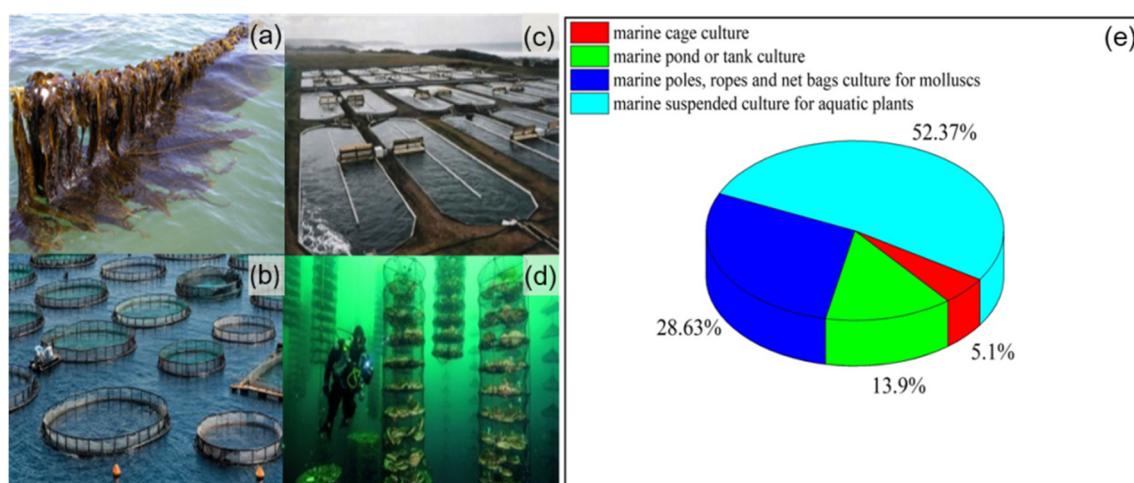


Figure 1. Types of typical mariculture systems: (a) marine suspended culture for aquatic plants; (b) marine cage culture; (c) marine pond culture; (d) marine poles, ropes, and net bags culture for mollusks; (e) the proportion of yield in typical mariculture systems (Food and Agriculture Organization of the United Nations, FAO, 2019 [27]).

3. Mariculture Effluents: Composition, Waste-Load, and Environmental Impacts

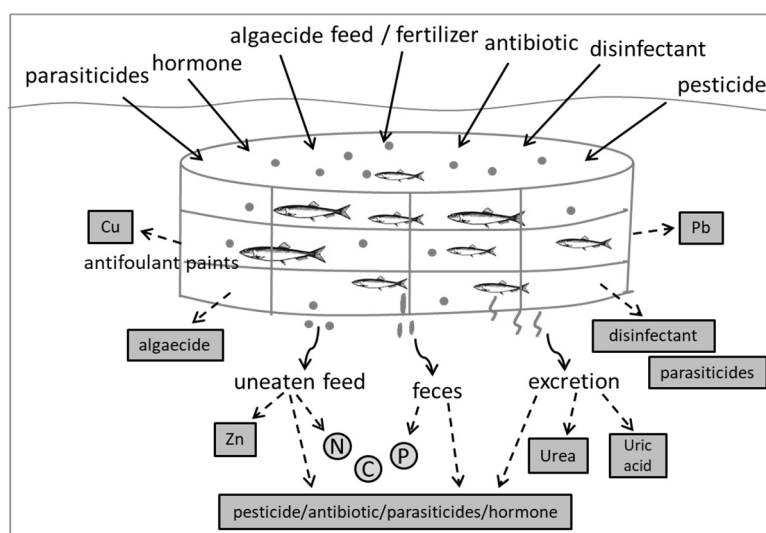
3.1. Composition

3.1.1. Organic Matter and Nutrients

The types of waste contained in the mariculture effluent mainly include organic matters such as organic carbon, urea, uric acid, and nutrients such as nitrogen (N) and phosphorus (P) [28]. These organic compounds and nutrients originate from fertilizers, feeds, and metabolic processes, such as uneaten feed, ingested but undigested food (feces), or food ingested and eliminated as excretion during production (Table 1, Figure 2). Fertilizers and feeds are introduced into cage or pond culture to promote the production of fish, shrimp, or other species. These substances contain high contents of protein, carbohydrates, lipid, nitrogen, and phosphorous, which could be transformed to dissolved organic matters and nutrients when in contact with water [29]. Detritus, uneaten feed pellets, and metabolic fecal matter provide organic particulates and nutrients in the water column, which can subsequently settle to, and deposit upon, seabed sediments [30].

Table 1. List of composition and main source of mariculture effluent in cage system.

Composition	Main Source	
Organic matter and nutrients	Organic carbon	Uneaten feed, feces
	Urea	Excretion
	Uric acid	Excretion
	Nitrogen (N)	Uneaten feed, feces
	Phosphorus (P)	Uneaten feed, feces
Chemicals	Antibiotic	Uneaten feed, feces, and biliary excretion
	Pesticide	Uneaten feed, feces, and biliary excretion
	Parasiticides	Uneaten feed, feces, biliary excretion, and water column (bath)
	Algaecide	Water column (bath)
	Disinfectant	Water column (bath)
	Hormone	Uneaten feed, feces, and biliary excretion
	Zinc	Feed, algaecides and pesticides
	Copper	Antifoulant paints, algaecides, and pesticides
	Lead	Antifoulant paints

**Figure 2.** Schematic diagram of mariculture effluent discharge.

3.1.2. Chemicals

Chemical contamination from the mariculture systems mainly include chemicals and heavy metals. Therapeutants are the most commonly used chemicals for a wide variety of purposes, including antibiotics to control disease, algaecides, pesticides, and parasiticides to control algae; pest species and parasites (e.g., sea lice) [31]; hormones to enhance productivity; and disinfectants to maintain hygiene throughout the production cycle (Table 1, Figure 2) [32]. Uneaten feed, feces, and biliary excretion are the main sources of antibiotics, pesticides, parasiticides, and hormones, whereas disinfectants and algaecides are commonly administered as a bath directly added to the water column and discharged outside the farm [33]. Moreover, heavy metals such as zinc, copper, and lead are often contained in antifoulant paints and coatings for the cage structures [34], ingredients of fish feed, or algaecides and pesticides such as copper sulfate to prevent infections and algal overgrowth [35]. These heavy metals can therefore find their way into marine environments with relatively high concentrations of mariculture infrastructure and effluents, and can concentrate both in the water column and benthic sediments in the area surrounding the cultivation site [36].

3.2. Waste Load

Waste load from mariculture systems has been largely studied with the main focus on nutrients discharge, especially N and P. Nutrient loading depends on several factors including the feed type

used, the proportion unconsumed, and fecal production [37]. The amount of feed required for cage cultivation is dictated by feed types, feeding frequency, and cage stocking density [38]. The actual amount of feed supplied to the cage that is consumed by fish and its digestibility are the two most important factors that determine the amount of uneaten feed and the fecal wastes produced [39]. In most studies, the nutrients released to the environment are calculated theoretically using models based on mass balance equations [40]. A traditional model to calculate the load of N and P uses a feed conversion ratio (FCR) based regression model that takes account of feed loss, FCR, and nutrient contents in the feed to predict the nutrient load for a given diet [41]. The model has been upgraded using the dry matter conversion rate (DMCR) instead of FCR to simulate more effectively the relationship between feed input, fish production, and waste generation [42]. Moreover, the nutrient load model can be coupled with fish growth models to simulate the temporal pattern of nutrient loads and identify maximum discharge during fish growth [43].

Chemical loading also depends on the types of chemicals utilized in treatments and the treatment frequency. Available chemical treatments in mariculture are strongly restricted by regulatory legislation and policies and may vary among different types of mariculture farms and between different countries. Discharge of chemicals such as medicines or antibiotics that are added directly to feed or fertilizer depends largely on the absorption rate of the fish and their fecal production [31,44]. In general, chemical load modeling has received relatively little attention, with conventional chemical assay mostly used to analyze chemical constituents in the effluent [45].

3.3. Environmental Impacts

The most significant negative environmental impacts of intensive mariculture are those related to high loads of organic matter, including nutrients and solids. Organic matter contributes to nutrient pollution, which further leads to hypoxia and eutrophication through fertilized aquatic plant growth and respiration [46]. Significant concentrations of organic matter and nutrients cause the excessive growth of phytoplankton, which reduces the transparency of the water column, affecting the photosynthesis of aquatic plants and impairing the balance between primary and secondary productivity [29]. Furthermore, eutrophication and high levels of organic matter lead to low dissolved oxygen concentrations in the water column. This may result in a reduction of wild fish populations and have detrimental impacts on other aquatic organisms as a result of deteriorated habitat quality [30]. Conversely, uneaten feeds can also provide a potential trophic resource used by biofouling communities attached to the cage structures, as well as wild fish in the surrounding waters. This can result in the aggregation of fish or other trophic creatures around the cage facilities, stimulating biological activity [47]. Organic and nutrient enrichment may also change the physicochemical compositions of seabed sediments and further affect the abundance, biomass and biodiversity of the benthic communities [48]. An increase of total organic carbon accumulation in the sediment would be harmful to benthos due to decreasing oxygen availability [49]. On the contrary, the enrichment in lipid content due to the accumulation of uneaten food pellets on seabed would become nutrient resources to promote microbenthic algal growth [8]. Besides, the environmental impact of effluent on benthic domain varies with different substrates and the soft substrates without vegetation tend to accumulate more pollution than vegetated areas [8].

Mariculture effluent can also increase the concentration of heavy metals and chemical compounds in the adjacent water column and seabed sediments. Heavy metals are accumulated in benthic sediments, where they are incorporated into several biological and chemical cycles [50]. High concentrations of zinc, copper, or other metal elements are toxic to macrobenthic fauna and microalgae, which are crucial elements in the food-chain for natural fisheries production [51]. Moreover, the residues of medicines, antibiotics, and vaccines found in the effluent are common sources of anthropogenic chemical pollution derived from mariculture operations [52]. They may harm unintended species outside the mariculture systems, leading to antibiotic resistance and other toxicity side effects, resulting in cascading effects throughout the marine food web [53].

4. Transport and Transformation Processes of Mariculture Effluent

The transport and transformation processes dictate the spatio-temporal distribution and the final fate of effluent discharged from the mariculture systems, which in turn determine their effects on sensitive marine habitats and ecosystems. Therefore, improved understanding of the physicochemical and biological processes associated with the transport and fate of mariculture-derived effluents is crucial for the appropriate regulation and management of mariculture operations around the world.

4.1. Transport and Transformation Processes in the Water Column

Traditionally, the process of transport of an effluent is divided into three phases: near-field, mid-field, and far-field, all of which correspond to different spatial and temporal scales [54]. After the effluent has been discharged, the processes of initial mixing and turbulent diffusion/dilution take place in the water column. The initial dilution process is determined by hydrodynamic conditions at the discharge site and is typically accommodated by a regulatory mixing zone. The ambient flow field can also be altered significantly by the presence of the cage structures due to their partial blockage of the flow, especially in locations with clustering of large quantities of cages [55]. Subsequently, the dissolved and particulate organic matters and nutrients are transported and mixed as passive tracers in the ambient flow, and can affect not only the near field but also distant areas after being transported over large distances (Figure 3) [56]. Mixing of dissolved wastes, due to dispersion and diffusion processes, also varies with hydrodynamic conditions (e.g., winds, waves, tides, and currents), which collectively influence how far a waste plume will spread from the source [57]. Previous studies have suggested that the mixing distance of dissolved waste is dependent primarily on local current speed (and direction), water depth, and total effluent load from the mariculture site [58]. It has also been suggested that at low water depths, dissolved nutrients tend to attain higher concentrations in the water column and spread more rapidly in a short time, while in deep waters, they tend to release nutrients during the settling process and spread more slowly [59]. Hence, the nutrients would spread more slowly and last longer in deep water than in shallow water [60].

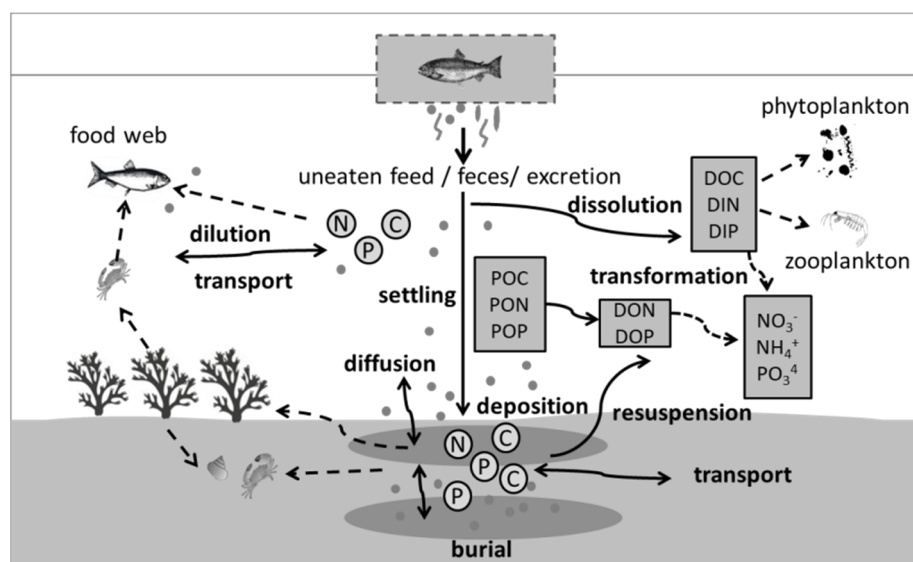


Figure 3. Schematic diagram of mariculture effluent transport and transformation processes.

The particulate fraction of mariculture wastes tend to settle in the water column after dilution. The settling velocity of the particles is mainly influenced by their physical characteristics such as density, the hydrodynamic conditions such as flow turbulence, and water viscosity [61]. For example, fecal particulates sink much more slowly than food pellets, partly because fecal particulates tend to absorb water and disintegrate, and thus reduce their density and settling velocity when descending

through the water column [56,62]. Flow turbulence is also critical to settling as it breaks particles into smaller ones and maintain them longer in suspension and thus need more time to settle [63]; however, its precise effect is still poorly understood at present. Moreover, environmental conditions such as the seawater temperature and salinity affect water viscosity and thus have indirect impacts on particulate settling velocities. Previous studies have shown that the settling velocities of particulate matter tend to be larger at higher temperature [61] and seawater salinity [64] due to the resultant reduction in water viscosity.

In terms of the transformation processes, as soon as uneaten feeds and fertilizers are discharged from the mariculture system, a proportion of dissolved organic matter and nutrients are diluted directly within the water column [65] and transformed into dissolved organic carbon [66], dissolved inorganic nitrogen (DIN) and phosphorus (DIP) (Figure 3) [29]. Detritus, uneaten feed pellets, and fecal particulates produced by metabolism can be transformed into particulate organic matter [67] and nutrients such as particulate organic carbon (POC), nitrogen (PON), and phosphorus (POP) present in the water column, which subsequently settles to the seabed and often form a deposited layer on the benthic sediments [30]. Dissolved organic nutrients (e.g., DON and DOP) can be further released through dissolution and resuspension from the deposited particulate fractions, while DIN, DIP, DON, and DOP present in the water column are further transformed into nitrates, ammonium, and phosphates through chemical processes such as ammonification, nitrification, and phosphorus assimilation [68]. The dissolution rate of particulate matter in the effluent is dependent on matter size, surface area–volume ratio, composition, and environmental conditions. It is known that smaller particulates with a larger surface area–volume ratios leach more nutrients for a given time, and nitrogen is dissolved at a faster rate than phosphorus, as the latter is mainly found in particulate form [69]. Besides, environmental conditions such as water temperature also affect the dissolution rate, which increases as temperature raises until the maximum dissolved concentration is reached [61]. Globally, the waste-load of nutrients from warm water systems would potentially be larger than from cold water systems under the same emission [70].

4.2. Transport and Transformation Processes at the Seabed Sediments

The larger particulates associated with mariculture effluents (e.g., feces, uneaten food) tend to gradually settle to the seabed and deposit on the sediment (Figure 3) [64]. The deposition of particulate wastes and spatial distribution on the seabed sediment depends primarily on the current conditions, particulate settling velocities and water depth, which collectively determine the overall settling period of the particles from their release from cages to deposition on the seabed. In quiescent waters, the maximum concentrations of deposited particulates are usually found directly beneath the fish cages, and their concentration declining gradually away from the cage area [71]. Moreover, previous studies have indicated that an increase in the current velocity appears to elongate the overall impact area of deposited particles, particularly on benthic sediments [72]. Meanwhile, deposited fecal matter tends to distribute more widely than food pellets on the seabed sediment under the same current, due to their lower settling velocity and greater sensitivity to the current velocity [60]. The deposition of individual particles is also attenuated by the bottom shear stress that low shear stress allows a larger probability for deposition [73].

When the deposited particulates have accumulated on the seabed over an extended period, they can be resuspended back into the water column when the threshold condition for resuspension is attained (Figure 3). The threshold condition for particulate resuspension from the seabed depends on the near-bed current velocity, shear stress, and water depth [74]. The critical bed shear stress is one of the main thresholds for resuspension, which depends strongly on substrate roughness and material type. Specifically, smooth substrates typically require lower critical stresses for material resuspension than rough surfaces [75], while the critical shear stress also varies greatly among different particulate materials with varying size and cohesiveness. For instance, smaller-sized particulates tend to have lower resuspension thresholds, and resistance to resuspension increases with increasing

cohesiveness, consolidation and biostabilization of the benthic sediment layer [63,76]. Moreover, numerical simulations have demonstrated that water depth also has a significant influence on the occurrence of particulate resuspension. For example, in a shallow water environment with a 12 m depth, one-third of the resuspension is attributed to surface waves, whereas at depths greater than 60 m, typically almost no particulate resuspension occurs and very limited mixing of deposited particulate wastes is observed [77].

The burial process of deposited particulates initiates when the deposited flux is larger than that of resuspension, which forms a deep sediment layer below the active deposited sediment layer [78]. The burial efficiency of particulate nutrients or organic carbon is found to be influenced by settling velocity and sediment porosity, and decreases with increasing depth [79]. However, the dominant controlling factors are not completely clear. Meanwhile, with continuous deposition and burial of particulate waste such as N and P at the seabed, a fraction of nutrients is diffused from porewater to the overlying water column. The diffusive flux is determined by concentration gradient of the sediment-water interface and effective diffusion coefficient in porewater, which is influenced by sediment tortuosity and porosity [80].

5. Regulatory Mixing Zones

5.1. Concepts of Regulatory Mixing Zones

Similar to marine wastewater discharges, the mixing zone for mariculture effluent is defined as the area immediately around the cage culture system where the effluent is physically mixed and dispersed into the surrounding environment [81]. Emission limit values (ELVs) and environmental quality standards (EQSs) are established in the mixing zone for important contaminants present in the effluent, which apply to the water quality at the end of point-source and within the receiving environment, respectively [82]. ELVs are set according to the assimilative capacity of the receiving waters to achieve the goal that the final environmental concentrations will comply with EQSs after mixing in the receiving waters [83]. Hence, a regulatory mixing zone (RMZ) is required to define a fraction of the MZ where EQSs can be exceeded, but higher environmental standards must be met outside the boundary of RMZ (Figure 4) [84]. The RMZ consists of the initial diffusion zone and the secondary mixing zone [85,86]: the initial diffusion zone is around the point-source where the initial contact with the receiving water occurs, and thus it is with the greatest concentration of the effluent [87], and the secondary mixing zone covers a more extensive area in which the effluent is mixed by turbulence so as to protect the whole waterbody from being polluted by the mariculture system [86]. Depending on the context, the RMZ is also termed as the allowable zone of effect (AZE) or inner zone, and some additional information regarding the RMZ is as follows [88–90]:

- (a) The RMZ applies only to the point-source discharge of effluent to a surface waterbody.
- (b) The RMZ contains the initial mixing process in a restricted area and is thus subject to the highest concentration of the effluent that could potentially exceed the EQSs therein. However, the water quality deterioration in the RMZ does not impair the integrity of the water body as a whole, and there is no lethality to organisms passing through the mixing zone.
- (c) The objective of the RMZ is to manage the load and water quality of effluent discharge and achieve the highest attainable level of ambient water quality or the minimum environmental impact from dilution and mixing.

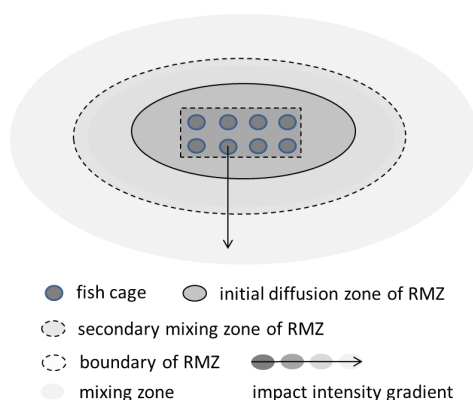


Figure 4. Schematic diagram of the regulatory mixing zone (adapted from Scottish Environment Protection Agency, SEPA [84]).

5.2. Characteristics of Regulatory Mixing Zones

The RMZ determines the extent to which effluent is allowed to initially mix and plays a vital role in limiting the pollution of receiving water bodies from the mariculture system. If the regulatory mixing zones are not defined and implemented appropriately, they could adversely affect mobile species passing through the mixing zone as well as settled species such as macrobenthic communities in the benthic sediments near the effluent sources [86]. The key characteristics and descriptions of RMZ in the major mariculture countries and regions, including location, size, and shape, are listed in Table 2. The member states of the European Union, including the accession countries such as Norway and the UK (including Scotland) [83], need to comply with the Water Framework Directive. In Australia and the United States [91], most states have their own legislation and policies but still broadly follow a nationwide management framework, and Canada tends to follow US EPA guidance for RMZ management [86]. Some major mariculture countries such as China are found to lack specific regulations on the designation of regulatory mixing zones as a prominent issue [92,93]. Amongst the various countries and regions, the RMZ is typically set concerning the location of the aquaculture system, while considering local ecosystem requirements and existing uses of the waterbody. The size of the RMZ appears to be more variable as a site-specific feature but mostly varying in the range of 100–200 m in radius. The shape of the RMZ is often not specified in the regulations of many countries.

Table 2. List of the key characteristics of regulatory mixing zones (RMZs) in different countries and regions.

Country/Region	Location	Size	Shape
EU (including Norway)	The mixing zone should be adjacent to points of discharge and up to 25 m from the cages [83].	The size should be as small as possible, without compromising the receiving water standards or significantly affecting aquatic life [83].	N/A
UK (including Scotland)	The mixing zone should lie within 100 m of the pens in all directions [84].	The size of the mixing zone should not exceed the equivalent symmetrical area [84].	The shape of the mixing zone does not have to be symmetrical [84]
USA and Canada	The mixing zone should be 100 m from the point of discharge [86].	The size should be as small as possible [86].	The shape should be a simple configuration that is easy to locate in a body of water [86].
Australia	The mixing zone is set up at 200 m from the mean low water mark of the coast at spring tides [91].	The size should not have a radius exceeding 100 m [91].	N/A
New Zealand	The mixing zone is limited to 200 m downstream of the point of discharge [89].	The size of the mixing zone will vary with the effluent flow, effluent concentration, background concentration, and bed characteristics [89]	N/A

6. Environmental Standards of Mariculture Effluent

In the context of regulation, regulators are required to establish the relevant environmental standards that apply to the quality of the effluent discharged at the end of point-source (ELVs) and receiving environment outside the regulatory mixing zones (EQSs), respectively, thus avoiding irreversible environmental impacts. Regular monitoring is also required to assess whether individual farms meet the required standards.

6.1. Emission Limit Values (ELVs) of Mariculture Effluent

ELVs of discharged effluent can be found in national and international legislation for different types of effluent discharge, e.g., the ANZECC (2000) water quality guidelines in Australia [91], the Federal Water Pollution Control Act (FWPC) (1972) in the USA [94], and the Marine Aquaculture Water Discharge Requirements (2007) in China [95]. The most relevant water quality parameters for mariculture effluent, including the abiotic and biotic parameters, are listed in Table 3. Some parameters are related to the discharge of organic matter and nutrients such as the concentration of dissolved oxygen (DO), biochemical oxygen demand (BOD), chemical oxygen demand (COD), total suspended solids (TSS), total nitrogen [96], total phosphorus [97], ammonia, nitrate, total organic carbon [98], and chlorophyll-a. Other parameters are related to the release of chemicals such as the concentration of copper, zinc, chromium, nickel, cadmium, lead, mercury, and free residual chlorine [95]. In addition, some fundamental features of the seawater such as pH and temperature are also included [99]. Notably, it is beyond the scope of this paper to provide an exhaustive overview of ELVs of mariculture effluent at a global scale, and the typical parameters in some major countries are documented in Table 3 for reference.

Table 3. List of abiotic and biotic parameters of emission limit values (ELVs) in different countries and regions.

Abiotic Parameters	Critical Concentration	Country	Reference
Free residual chlorine	<0.1 mg/L	China	[95]
	<13 µg/L	USA	[86]
	4–6 mg/L	Canada	[99]
Dissolved oxygen	>6 mg/L	China, Australia	[95,100]
	>5.7 mg/L	UK	[101]
	<10 mg/L	Australia	[100]
Turbidity	<12 mg/L	UK	[101]
	≤10 mg/L	China	[95]
Chemical oxygen demand	≤6 mg/L	China	[95]
Biochemical oxygen demand	<10 mg/L	Australia	[100]
	6.5–9.0	Australia	[100]
	6.5–8.5	USA	[86]
pH	6.0–9.0	Canada	[99]
	7.0–8.5	China	[95]
Seawater temperature	the optimum temperature for shrimp is 27 °C	Canada	[99]
Total suspended solids	<100 mg/L	Canada	[99]
	<40 mg/L	China	[95]
Total nitrogen	<3 mg/L	Australia	[87]
Total phosphorus	<0.5 mg/L	Australia	[100]
Ammonia	<2 mg/L	Canada	[99]
Nitrate	<0.09 mg/L	Canada	[99]
Total organic carbon	<10 mg/L	Australia	[100]
CaCO ₃	>100 mg/L	Canada	[99]
Sulfide	<0.002 mg/L	Australia	[100]
	<4.8 µg/L	USA	[86]
	<5 µg/L	China	[95]
Copper	<3.76 µg/L	UK	[101]
	<90 µg/L	USA	[86]
	<20 µg/L	China	[95]
Zinc	<5 µg/L	Australia	[100]
	<6.8 µg/L	UK	[101]
	<1100 µg/L	USA	[86]
Chromium	<74 µg/L	USA	[86]
	<100 µg/L	Australia	[100]
Nickel	<40 µg/L	USA	[86]
Cadmium	<210 µg/L	USA	[86]
Lead	<0.05 µg/L	Australia	[100]
Mercury			
Biotic Parameters	Critical Concentration	Country	Reference
Chlorophyll-a	0.01 mg/L	Canada	[99]
Total marine bacteria	1000 CFU/L	Canada	[99]

6.2. Environmental Quality Standards (EQSs) of Regulatory Mixing Zones (Receiving Environment)

Many countries have proposed environmental quality standards and regulatory policies to safeguard the water and sediment quality of receiving environment, e.g., the Marine Water Quality Standard (1997) in China [102], the National Aquaculture Policy Framework (2013) in South Africa [103], the Aquaculture Management and the Environment Protection (Water Quality) Policy (2003) in Australia [104], the Technical Support Document for Water Quality-based Toxics Control (USEPA 1991), and Water Quality Standards Handbook: Second Edition (USEPA 1994) in the USA [105]. The EQSs are prescribed with respect to the RMZ, and mainly concern water quality and sediment quality.

6.2.1. Water Quality Standards

The key environmental standards associated with the receiving waters of mariculture effluents are associated with water quality within the RMZ and the fact that this should not impair the integrity of the receiving water as a whole, and that there is no lethality to organisms passing through the RMZ. Therefore, many countries and regions have proposed environmental standards to manage the water quality of the RMZ (Table 4). Generally, the parameters are part of those applied to effluent discharged, particularly the abiotic parameters such as suspended solids, total nitrogen, total phosphorus, and dissolved oxygen.

Table 4. List of water quality standards of regulatory mixing zone (receiving environment) in different countries and regions.

Country	Standards of Water Quality	Reference
China	Dissolved oxygen > 6 mg/L; Chemical oxygen demand ≤ 1 mg/L	[102]
Australia	Concentrations of suspended solids, total nitrogen, and total phosphorus must not exceed background concentrations (determined by administering authority) at a boundary of the regulatory mixing zone.	[87]
USA	Pollutant concentrations are not lethal to organisms passing through the mixing zone.	[86]
UK	Concentrations of organic waste outside of the mixing zone are low enough to avoid adverse impacts on the status of sea life.	[84]
South Africa	Dissolved oxygen ≥ 60%.	[10]

6.2.2. Sediment Quality Standards

Many countries and regions have also proposed environment standards to regulate the sediment quality of the RMZ (Table 5). The parameters include biotic ones such as contents of sulfide, carbon, zinc, copper [106], and biotic ones, such as biodiversity index and composition of benthos [107]. It is evident that considerable variability still exists in the parameters and the respective critical concentrations prescribed from different countries, which need to be thoroughly examined in the future to achieve better consistency across the world.

Table 5. List of sediment quality standards of regulatory mixing zone (receiving environment) in different countries and regions.

Country	Standards of Sediment Quality	Reference
USA	Sulfide ≤ 1500 µM; the Shannon–Wiener diversity index of benthic < 0.5 and percent Capitella capitata of benthic > 25%.	[107]
UK	Zinc ≤ 410 µg/g, copper ≤ 270 µg/g, indicating benthic impacts that are “probably adverse”, and Zinc ≤ 270 µg/g, copper ≤ 108 µg/g represent benthic impacts that are “potentially problematic”.	[106]
Norway	Total organic carbon ≤ 20 mg/g.	[108]
South Africa	Faunal benthic composition, electroconductibility, sediment carbon content.	[10]

6.3. Techniques for Environmental Monitoring

Once environmental quality standards are established, monitoring measures to assess the mariculture compliance is also required to be put in place, with the purpose of acquiring and analyzing environmental data at various spatio-temporal scales to evaluate whether the effluent discharge meets the required environmental standards [10]. If the concentration and/or impact area is found through monitoring to be larger than the maximum allowable standards (i.e., defined by the regulatory mixing zone (RMZ) and allowable zone of effect (AZE)), appropriate regulatory steps are taken to reduce the effluent discharge and thus the resultant environmental and ecological impacts. In recent years, mariculture effluent monitoring techniques have developed rapidly from traditional manual point sampling techniques (e.g., visual diver surveys [109]) to cover increased monitoring areas using more time-efficient and cost-effective approaches (e.g., remote sensing [110], video monitoring [111]). In the following, we mainly focus on current mainstream monitoring techniques that permit both wide spatial coverage and high temporal resolution in data collection of important physical and biochemical parameters.

Wireless sensor networks (WSNs) have been recently adopted to provide continuous monitoring of water quality at high spatial and temporal resolutions. These systems allow a large number of water quality sensors to be installed around the mariculture site for detailed monitoring of dissolved oxygen (DO), salinity, total suspended solids (TSS), temperature, water level, and ammonia nitrogen [112,113]. Compared with more traditional point sampling techniques, WSNs are more reliable and cover the widespread dispersal of the effluent plume [114]. The application of remote sensing in mariculture monitoring is also now widely conducted using an array of optical (e.g., Landsat Thematic Mapper (TM), China–Brazil Earth Resources Satellite (CBERS)-3/4 and Sentinel-2A), and radar-based space-borne instruments (e.g., CosmoSkyMed, Radarsat and Sentinel-1A), to measure and monitor water quality parameters such as chlorophyll-a, suspended particulate matter, turbidity, and colored dissolved organic matter at varying spatio-temporal scales [110]. Acoustic doppler current profilers (ADCP) combined with free-fall multi-parameter profiler could be used to monitor the flow structures and water quality parameters such as conductivity, temperature, fluorescence, turbidity, and dissolved oxygen to show the possible distribution of effluent materials [115]. In addition, in-situ video surveys are used to monitor changes to benthic communities resulting from mariculture without the need for destructive sampling [111], as well as the feed uptake by fish stocks within the mariculture cages themselves.

Complementary to physical techniques, biochemical monitoring can trace the source of waste from aquaculture system to further identify environmental change caused by aquaculture effluent. Biochemical tracing techniques such as fatty acid (FA) analysis [116], stable isotopic signatures [117] and soya DNA [118] are relatively recent innovations now used increasingly as effective monitoring tools to trace mariculture waste in the water column or benthic system. They could effectively differentiate between aquaculture waste and marine source waste, as well as information on the overall influence of mariculture in marine ecosystems. Moreover, biological indicators are also widely used to monitor effluent discharge and the quality of receiving environment. Amongst them, macroalgal bioindicators can monitor and identify nutrient pollution between multiple effluent sources in the water column [119], whereas protistan groups can be used as bioindicators in environmental DNA barcoding methods to monitor the environmental impacts on macrobenthic communities [120].

7. Water Quality Modelling of Mariculture Discharges

The objective of achieving sustainable mariculture production hinges on the knowledge of the local and regional interactions between mariculture effluents and marine ecosystems. It is thus crucial to be able to predict the dispersal, deposition, resuspension, and accumulation of particulate organic matter, as well as dissolved nutrients and chemicals that emanate from the mariculture site within the surrounding near-field (local) and far-field (regional) pelagic and benthic habitats. Additionally, we need to accurately predict forecasting for the carrying capacity of the mariculture site, the assimilate capacity of receiving water body, and the severity of potential environmental impacts [121]. A number

of modelling techniques have been developed to simulate the magnitude and extent of potential contamination from mariculture activities (Table 6), informing both the regulatory framework under which such activities are licensed and the site management and monitoring activities to ensure that compliance is achieved.

Table 6. List of models used in the management of mariculture effluents.

Model	Simulated Systems/Dimensionality	Model Structure	Simulated Processes	Applications
Near-Field Model				
DUCHESS [122]	<ul style="list-style-type: none"> water column 2-dimensional 	hydrodynamics model coupled with particle tracking model	particle waste transport process	aquaculture particle waste transport tracking [123]
NewDEPOMOD [124]	water column and seabed sediment	particle tracking model integration with regional hydrodynamic models (e.g., FVCOM)	deposition of waste to the seabed	used in the site selection of a new farm to investigate the proposed farm position and biomass levels [125]
MERAMOD [126]	water column and seabed sediment	particle tracking model coupled with benthic response modules	deposition of waste to the seabed	assess the benthic effects of particulate waste in Mediterranean area [13]
CODMOD [127]	water column and seabed sediment	particle tracking model coupled with cod-specific data.	dispersion and resuspension process of waste feed and fecal particles	assess the benthic effects of cod farm [128]
Far-Field Model				
Aquaculture waste transport simulator (AWAST) [129]	<ul style="list-style-type: none"> water column 2-dimensional 	flow model coupled with particle tracking model	physical dispersion of finfish aquaculture wastes	safeguard the long-term use of fish farm sites [123]
Mike21 [130]	<ul style="list-style-type: none"> water column 2-dimensional 	MIKE21 HD, MIKE21 AD, MIKE21 PT	<ul style="list-style-type: none"> spreading of dissolved matters transport and dispersal of suspended feces 	<ul style="list-style-type: none"> optimization of mariculture systems environmental impact assessment of mariculture infrastructures [41]
Bergen Ocean Model [74,131]	<ul style="list-style-type: none"> water column and seabed sediment 3-dimensional 	hydrodynamic σ -coordinate model coupled with a Lagrangian particle tracking simulator	dispersion and deposition process of particulate organic matter	particulate organic matter dispersion and deposition simulation [132]
Marine Environmental Committee (MEC) ocean model [133]	<ul style="list-style-type: none"> water column 3-dimensional 	coupled with submodules of fish cage drag and the diffusion of particulate organic waste.	particle waste transport process	provide references for the time and location of IMTA systems [134]
TELEMAC-3D [135]	<ul style="list-style-type: none"> water column 3-dimensional 	hydrodynamics model coupled with particle tracking model	dispersion process of particle wastes	assess the installation of new aquaculture systems [59]
MOHID [136]	<ul style="list-style-type: none"> water column 3-dimensional 	hydrodynamics model coupled with particle tracking model	dispersion process of particle wastes	simulate the particles dispersion through Lagrangian tracers in the far-field [137]
Delft3D [138]	<ul style="list-style-type: none"> water column 3-dimensional 	hydrodynamics model coupled with particle tracking model	dispersion process of particle wastes	simulate the waste dispersion in the offshore area and provide a reference for site selection [139]
LAMP3D [140]	<ul style="list-style-type: none"> water column 3-dimensional 	hydrodynamics model coupled with particle tracking model	transport of dissolved nutrients, fecal particulates and uneaten feed	simulate the time-varying dispersion with different pollutants and/or with different conditions of release [141]
Finite-Volume Community Ocean Model (FVCOM) [142]	<ul style="list-style-type: none"> water column and seabed sediment 3-dimensional 	hydrodynamic model coupled with sediment transport model	transport, deposition and resuspension of particulate wastes	predict the far-field distribution and potential environmental effects of particulate waste [143]
Regional Ocean Modelling System (ROMS) [144]	<ul style="list-style-type: none"> water column and seabed sediment 3-dimensional 	hydrodynamic model coupled with particle tracking model	transport, and deposition of particulate wastes	simulate the time-varying impact on benthic sediment with changes in fish feeding behavior [141]
Coupled Near- and Far-Field Model				
SUNTANS [145]	<ul style="list-style-type: none"> water column 3-dimensional 	two-dimensional shallow water equations together with the depth-averaged continuity equation	mixing and transport process of waste under oscillatory flow conditions	assess the impact of aquaculture fish pens on coastal water quality in both near- and far- field

Table 6. Cont.

Model	Simulated Systems/Dimensionality	Model Structure	Simulated Processes	Applications
Ecological Carrying Capacity Assessment Model				
Modelling ongrowing fish farms monitoring (MOM) [146]	system-scale	fish model, coupled with dispersion model and benthic model	<ul style="list-style-type: none"> release of particulate material from the farm dispersion and deposition of organic waste decomposition at the benthic boundary layer 	provide references for the size of fish farms and the hydrodynamic and topographic conditions of the selected areas [147]
SMILE framework [148]	system-scale	Delft3D model coupled with cultured species growth model (ShellSIM) and ecosystem models (EcoWin2000)	interactions between environmental variables and aquaculture activities	<ul style="list-style-type: none"> estimate the carrying capacity in shellfish farms tools for management in sustainable development of shellfish aquaculture [149]
Environmental Impact Assessment Model				
Integrated Coastal Zone Management (ICZM) [150]	system-scale	3-level nested hydrodynamic model coupled with particle tracking module and GIS model	dispersion and deposition of organic waste	The model offers an open source alternative to particulate waste modeling and evaluation, contributing valuable information for local decision makers in the process of locating new facilities and monitoring stations. Offers information for new farm and monitoring stations locating [151] assess the ecological risk of therapeutants from aquaculture effluent [153]
Site-specific multimedia fate model [152]	system-scale	JETLAG model integrated into the fugacity model	transport and transformation of chemicals	

7.1. Near-Field Models

Water quality models simulating the near-field impacts from mariculture sites have been applied in many areas such as Scotland, Norway, and the Mediterranean [60]. Amongst them, a simple two-dimensional model termed DUCHESS has been developed to simulate particle waste dispersion from cages in the water column [123]. Models for particle settling, deposition, resuspension, and transport have also been developed to assess the impacts of particulate wastes on the benthic ecosystem. Simple particle tracking models such as AutoDEPOMOD [125] have been used to predict the deposition of waste material to the seabed without considering the prevalent hydrodynamic conditions in the receiving water body. The model predicted the benthic impacts from a fish farm site on a 1 km-by-1 km area around the farm [154]. Subsequently, mariculture waste deposition models such as MERAMOD have been developed to track organic solids particle incorporation and degradation in seabed sediments by combining the hydrodynamic model with a transport module, and predict the response of the benthic community [126]. The near-field resuspension model such as CODMOD that predicts the subsequent resuspension of bed materials in cod farming has been further developed by combining the particle tracking model with the resuspension model and incorporating particulate resuspension thresholds [127,128]. More recently, the NewDEPOMOD has been further developed to incorporate the resuspension process and the effects of varying bathymetry in the deposition process, along with potential integration with regional hydrodynamic models (e.g., FVCOM) to simulate more realistic flow patterns [124] and thus provide a link to far-field impact modelling.

7.2. Far-Field Models

Recent modelling efforts have been shifted gradually towards assessing the far-field effects of mariculture effluent that can extend kilometers away from individual sites, as well as investigating and quantifying the potential interactions between adjacent farms. These far-field models differ considerably from near-field modelling approaches in that they predict the effluent dispersion under varying hydrodynamic conditions at regional scales [7], while some far-field water quality models simulate the waste transport process in the water column. For example, the aquaculture waste transport simulator

(AWATS) was a two-dimensional flow model coupled with a particle tracking model to simulate organic waste transport under spatially varying current around the study area [123]. The Mike21 hydrodynamic model is also a 2D model that has been used to simulate the spreading of dissolved wastes and particle feces through the advection and dispersion processes [41]. Subsequently, many hydrodynamic models have been used widely by coupling them to other water quality models to simulate the dispersion of mariculture effluent. For example, the MEC (Marine Environmental Committee) hydrodynamic model has been coupled with a fish cage drag model and particulate organic waste transport model to simulate the far-field impacts from mariculture effluents [134]. Other three-dimensional hydrodynamic models, such as the Bergen Ocean Model [74,132], TELEMAC-3D [59], MOHID [137], and Delft3D [139], have also been coupled with particle tracking models to simulate particle dispersion in the far-field. Another recent model development is the LAMP3D model that couples a fully 3D hydrodynamic model (the Princeton Ocean Model, POM) and a Lagrangian transport model to simulate the time-varying mixing and impact of the solid wastes, in which the changes in fish feeding behavior has also been considered [141]. Other models also simulate the far-field impact of settled particles deposition, transport and impact on the benthic sediment. Some hydrodynamic models, such as the Finite-Volume Community Ocean Model (FVCOM) coupled with a 3D Lagrangian particle tracking model, have been used to simulate the far-field impact of feed solid waste transport and deposition on sediment and benthic ecosystems [143]. Similarly, the Regional Ocean Modelling System (ROMS), also coupled with the Lagrangian transport model, was further developed to simulate the time-varying impact of the solid wastes on benthic sediment with changes in fish feeding behavior [141].

7.3. Coupled Near- and Far-Field Model

Notably, there are also some models that simulate the impact of mariculture waste dispersion and transport in both near- and far-field. For instance, SUNTANS incorporates the local/regional seabed environment (e.g., bathymetry, sediments) and prevalent hydrodynamic conditions (i.e., tides and wind-induced currents and waves) to assess the waste transport behavior in both near-field fish-farm site and far-field regional area [57].

7.4. Modelling Techniques for Ecological Carrying Capacity and Environmental Impact Assessment

Models have been developed to simulate the ecological carrying capacity of the mariculture system and are used as effective management tools in site selection and coastal zone management [155]. Some models used to assess the ecological carrying capacity have also incorporated the process of effluent discharge. For example, the management system model MOM (modelling on-growing fish farms monitoring) has been widely used to analyze the carrying capacity of cage culture, which couples the fish model, water quality model, and benthic model, assisting in maintaining satisfactory environmental conditions around fish farms [147]. The SMILE framework has developed a comprehensive carrying capacity model by coupling the Delft3D model with cultured species growth model (ShellSIM) and ecosystem models (EcoWin2000) to assess aquaculture stocks of shellfish over multi-year periods [149].

Yet some other models have been developed to assess the environmental impacts of cage mariculture on marine ecosystems, in which effluent discharge modules also have been incorporated to predict waste dispersion impacts. For example, the Integrated Coastal Zone Management (ICZM) tool has been established by coupling a three-level nested hydrodynamic model, particle tracking module and GIS model to assist in locating new facilities and monitoring stations [151]. The site-specific multimedia fate model has also been developed and used to assess the environmental risk of chemical waste (e.g., antibiotics) as well as the spatial distribution of chemical concentration in the water column [153].

Generally, state-of-the-art water quality models have been applied to simulate the near-field impact of effluent wastes on water column and benthic sediment around fish farms, as well as far-field impact on local or regional marine system. In addition to the modelling of dissolved waste transport, some models are also integrated with a particle tracking module, to predict the transport of particulate

waste. However, modeling of the nutrients such as carbon and phosphorus cycle is still lacking, as well as a proper consideration of food web effects through incorporation of essential trophic levels such as phytoplankton. At the same time, the applications of far-field models are still subject to high computational cost and instability issues, and are less adaptive to discharge management and regulatory compliance. Finally, an efficient and effective coupling of near- and far-field models poses another prominent challenge as for other marine wastewater discharge modelling practices.

8. Conclusions & Recommendations

Effluent discharges from mariculture sites, especially cage cultures, are mainly composed of organic matter, nutrients, and chemicals from the uneaten feed, feces, and excretions, which may pose negative effects on the marine environment and ecosystems. In this study, we have reviewed the characteristics of effluents, their transport and transformation processes, as well as the associated environmental impacts on the water column and seabed sediments. Second, we have compared the relevant environmental standards of mariculture effluent including the concept of regulatory mixing zone from major mariculture countries. The monitoring techniques and water quality models most widely applied to assess the environmental impacts of mariculture effluents on near- or far-fields have also been reviewed to identify the state-of-the-art techniques and need for further development. It is found that the waste loading fluxes and transport processes of effluent depends primarily on feeding activities, stocking capacity, and environment conditions. The objective of the regulatory mixing zone is to minimize the environmental impacts of effluent through dilution and mixing. Emission limit values and environmental quality standards are established in many countries to regulate the discharged effluent quality. In addition, regular monitoring is also required to assess whether actual farming activities meet the required standards. Considerable variability has been found in the effluent parameters and the respective critical contaminant concentrations prescribed from different countries, which need to be thoroughly reviewed in the future.

Many of the main countries involved in mariculture activities, such as China, currently lag behind in the management and regulation of effluents (e.g., specific regulations are lacking in the location, size, and environmental standards of mixing zones) compared to Western Countries, which therefore needs to be addressed to promote sustainable development within the mariculture industry. Several water quality models have been developed in the past decades as cost-effective tools to study and predict transport and transformation processes of mariculture effluents and the related environmental impacts as well as, ultimately, for mariculture regulation. However, a proper incorporation of the ecological impacts such as the food web effects as well as the effective coupling of near- and far-field models are some of the open challenges needed to be addressed. Whilst new models are continually emerging, regulatory authorities are increasingly turning to predictive models to make informed decisions regarding the licensing of new marine fish farms with their effluents discharged into receiving waters. As such, the need to evaluate and validate these predictive models remains imperative to ensure their appropriate use and application in the management of mariculture sites. Monitoring projects are also continuing to play a vital role in mariculture planning and management and can be optimized by combining with water quality models to enhance efficiency and effectiveness.

Overall, this paper is intended to provide a state-of-the-art review of characterization and management tools of mariculture effluents in order to push forward their regulatory management and, ultimately, to promote a sustainable mariculture development into the future.

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